RESEARCH

Open Access



The association of test diameter and gonad weight with some toxic trace metals level in black sea urchin (*Stomopneustes variolaris*)

H. K. S. De Zoysa^{1,2*}, B. K. K. K. Jinadasa³, E. M. R. K. B. Edirisinghe² and G. D. T. M. Jayasinghe³

Abstract

Background: Quantification of toxic trace metals in biota is very important to monitor both ecosystem health and public health. *Stomopneustes variolaris* is a widespread species of sea urchin in the world that have edible gonads that are in high demand. This study attempted to examine the levels of accumulated toxic trace metals and their relationship between the test diameter and the gonad weight of the black sea urchin (*S. variolaris*) in Sri Lanka.

Results: As *S. variolaris* is highly abundant in southwest coast of Sri Lanka, three sampling sites were selected in this region (Mount-Lavinia, Beruwala and Tangalle). The concentration of toxic trace metals (Hg, Cd, Pb and As) in the gonads of *S. variolaris* was determined in randomly selected subset from the above samples. The levels of above metals in gonads did not exceed the permitted standard levels of the EU and Sri Lanka. The mean size and weight of *S. variolaris* collected from Mount-Lavinia was 5.55 ± 1.04 cm and 101.40 ± 57.76 g, respectively; for Beruwala it was 6.54 ± 0.86 cm and 147.90 ± 50.40 g, respectively, and for Tangalle it was 6.41 ± 1.05 cm and 150.50 ± 59.45 g respectively. There was a significant relationship between all the analysed trace metals and the test diameter and gonad weight of *S. variolaris* in Sri Lanka.

Conclusions: This study revealed that the trace metal levels in gonads of *S. variolaris* did not exceed the harmful level for human consumption.

Keywords: Accumulation, Echinodermata, Roe, Indian Ocean

Background

Heavy metals that are less soluble in water tend to accumulate in marine ecosystems as they do not degrade [1]. Therefore, these toxic trace metals will accumulate along the trophic levels which will affect the entire marine community. The levels of trace metal accumulation in marine organisms depend on their detoxification mechanism in metabolism. The harmfulness of these toxic trace metals is due to their role in biochemical and metabolic processes. Excessive consumption of sea food with high levels of accumulated toxic trace metals can lead to

*Correspondence: dezoysahks@yahoo.com

¹ Department of Bioprocess Technology, Faculty of Technology, Rajarata University of Sri Lanka, Mihintale, Sri Lanka

Full list of author information is available at the end of the article



metabolic dysfunctions. Therefore, quantification of toxic trace metals in the all biota is very important to monitor both ecosystem health and public health. Currently, sea urchin gonads are a delicacy in many countries, and thus, there is a very high demand. However, nothing is known about the levels of accumulated toxic trace metals in sea urchins [1, 2].

The black sea urchin, *Stomopneustes variolaris*, is one of the edible species of sea urchin in the world. In Sri Lanka, *S. variolaris* is restricted to the western and southern coastal areas than other coastal areas [3, 4]. The black sea urchin (*S. variolaris*) is one of the warm water sea urchin species that belongs to the family Stomopneustidae. They bore into the rock, reefs and other substrates in littoral regions up to a depth of 18 m, and

© The Author(s) 2018. This article is distributed under the terms of the Creative Commons Attribution 4.0 International License (http://creativecommons.org/licenses/by/4.0/), which permits unrestricted use, distribution, and reproduction in any medium, provided you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The Creative Commons Public Domain Dedication waiver (http://creativecommons.org/ publicdomain/zero/1.0/) applies to the data made available in this article, unless otherwise stated.

they prefer to live in the shady areas. *S. variolaris* is an omnivore species, and also they are known to mainly feed on algae and seaweeds [5–13]. *S. variolaris* species has a widespread distribution [12, 14, 15], in tropical and subtropical regions of the Indo-Pacific from the east African coast and Madagascar to New Caledonia and Samoa to the Bonin Islands to the north [12]. It is an abundant edible species in the Indian Ocean, in the geographical regions such as Lakshadweep Islands, Andaman Islands, India and Sri Lanka [16]. The growth rate of sea urchins and natural mortality mainly depend on temperature and food availability [6]. The sea urchins accumulate nutrients in the gonad, which is a highly sought-after delicacy in the world [17–19].

The trace metals are accumulated in most body parts of echinoderms including coelomic fluid, spines, gut, gonads and Aristotle's lantern. Hence, levels of trace metals in sea urchins could be used as a bio-indicator to determine the level of environmental pollution [20]. Currently, nothing is known about the levels of trace metals in the gonads of sea urchins in Sri Lanka, which has high demand in the marine fisheries export industry [18, 19]. Therefore, this study was conducted to get an idea on the levels of toxic trace metals in *S. variolaris* in Sri Lanka.

Methods

A total of 197 individuals of S. variolaris were collected from three selected sites in the western and southern coasts of Sri Lanka, namely Mount-Lavinia (n=43), Beruwala (n=99) and Tangalle (n=55) reefs (Fig. 1) from May to November in 2014. All collected individuals were transported immediately with seawater to analytical chemistry laboratory at the Institute of Post-Harvest Technology (IPHT), National Aquatic Research Agency (NARA) within 24 h. After that, total body weight to nearest 0.01 g and the horizontal test diameter was measured to the nearest 0.02 mm using a vernier caliper. Horizontal test diameter was measured twice perpendicularly to the first measurement and the averages of the two measurements were taken. Finally, gonads were removed from each individuals and weighed after removing excess water using blotting papers.

To determine the toxic trace metal levels in gonads of *S. variolaris*, individuals with test/shell diameter between 3 and 9 cm were selected and 6, 13 and 11 individuals were selected from Mount-Lavinia, Beruwala and Tangalle, respectively [21]. The levels of toxic trace metals, Hg (mercury), Pb (lead), Cd (cadmium) and As (arsenic) were assayed in the atomic absorption spectrophotometer (AAS) (Varian 240 FS, Springvale, Australia). The standard solutions of Hg, Pb, Cd and As at 1000 mg/L (ppm) were obtained from Sigma-Aldrich (Dorset, United Kingdom) which were used to calibrate. All standards and reagents were prepared using ultrapure water. One gram of homogenized sea urchin gonad samples were taken into a Teflon microwave digestion tube and all the samples were performed in duplicates. Acid digestions of gonad samples were performed by using the 10 mL of HNO₃ (65% nitric acid, AR) allowed to 10–15 min and capped the microwave digestion tubes. The sea urchin gonad samples were digested using the CEM/MARS XP-1500+microwave oven (CEM, Matthews, USA). Microwave accelerated system and allowed to cool room temperature and pressure was released carefully by opening the valve. Then the digested gonad samples were transferred into 50 mL of volumetric flasks and volumed up using the de-ionized water was used to determine the Hg, Pb, Cd and As of digested sea urchins' gonads samples. A graphite tube atomizer (Varian GTA-120) was used to determine the Pb and Cd, for the determinations of As palladium nitrate $(Pd(NO_3)_2)$ was used as a modifier. The Hg concentration of sea urchins' gonads were determined by using the cold vapour techniques (Varian VGA 77) using the $SnCl_2$ as a reduction [3, 4, 22].

The Pearson's correlation coefficient (r) and significant correlation tests were conducted to assess the relationship between gonadal trace metals levels and test diameter and gonad weight of *S. variolaris*. Data were analysed using the Minitab 16.0 version and Microsoft Excel 2010 version at the level of significance at $\alpha = 0.05$ (5%).

Results

Toxic trace metals levels in gonads

Levels of toxic trace metals in gonads of *S. variolaris* were determined separately from selected sites and the average values of accumulated concentration (ppm and ppb) and the total amount of trace metals with their average test diameter and gonad weight in selected individuals' are determined.

The average concentration (ppb) of Hg in selected individuals from Mount-Lavinia reef (n=11) ranged from (minimum to maximum) 0.00 (±0.00) to 14.40 (±5.33) µg/kg and total concentration in gonads varied from 0.00 to 0.27 µg (Table 1). The trace metal Cd ranged from 47.81 (±19.58) to 336.10 (±21.46) µg/kg and total concentration in gonads varied from 0.08 to 6.28 µg. Pb was not detected in the selected individuals from Mount-Lavinia. However, concentration of As ranged from 0.00 (±0.00) to 1805.87 (±104.54) µg/kg, and total accumulation by gonads varied from 0.00 to 33.75 µg.

For the Beruwala reef population, accumulated toxic trace metals (n=13) are shown in Table 2. The average concentration of Hg in the Beruwala reef population ranged from 34.40 (±1.62) to 104.39 (±2.18) µg/kg, and the total level in the gonads varied from 0.07 to 0.44 µg. Cd levels ranged from 150.72 (±1.81) to 1149.55



 $(\pm 90.07)~\mu g/kg$ and total levels in gonads varied from 0.17 to 9.86 $\mu g.$ Pb ranged from 0.00 (± 0.00) to 206.27 $(\pm 90.07)~\mu g/kg$ and total level in gonads varied from 0.00 to 0.36 $\mu g.$ Also As content in the selected individuals

ranged from 1142.52 $(\pm\,83.51)$ to 4635.66 $(\pm\,90.07)$ µg/ kg and total accumulation by gonads varied from 2.99 to 22.23 µg.

	•								
Test diameter	Total gonad	Hg		Cd		РЬ		As	
(cm)	weight (g)	Average (Hg) (µg/kg)	Total Hg (µg)	Average (Cd) (µg/kg)	Total Cd (µg)	Average (Pb) (µg/kg)	Total Pb (µg)	Average (As) (µg/kg)	Total As (μg)
3.70	1.77	0.00 (土 0.00)	0.00	47.81 (土 19.58)	0.08	0.00 (土 0.00)	0.00	0.00 (土 0.00)	0.00
4.50	1.92	00.00 (土 0.00)	0.00	48.38 (土 9.42)	0.09	0.00 (土 0.00)	0.00	1637.16 (土396.96)	3.14
5.10	3.10	9.68 (土0.82)	0.03	156.69 (土 2.63)	0.49	0.00 (土 0.00)	0.00	50.27 (土156.08)	0.16
6.20	10.07	00.00 (土 0.00)	0.00	158.03 (土 2.73)	1.59	0.00 (土 0.00)	0.00	305.31 (土64.61)	3.07
7.10	8.63	4.74 (土0.91)	0.04	217.00 (土 0.78)	1.87	0.00 (土 0.00)	0.00	434.28 (土324.83)	3.75
8.90	18.69	14.40 (土5.33)	0.27	336.10 (土 21.46)	6.28	0.00 (土 0.00)	0.00	1805.87 (土104.54)	33.75

Table 1 Toxic trace metals (Hg, Cd, Pb and As) concentration and total amount of particular trace metals present in *S. variolaris* at Mount-lavinia reef (*n*=6) with standard deviation (±5D)

WILL JU									
Test diameter	Total gonad	Hg		cd		Pb		As	
(cm)	weight (g)	Average (Hg) (µg/kg)	Total Hg (µg)	Average (Cd) (µg/kg)	Total Cd (µg)	Average (Pb) (µg/kg)	Total Pb (μg)	Average (As) (µg/kg)	Total As (μg)
4.90	1.15	61.91 (土 7.58)	0.07	150.72 (土1.81)	0.17	0.00 (± 0.00)	0.00	2600.20 (土 1.14)	2.99
5.30	2.13	54.38 (土 23.32)	0.12	254.07 (土 15.25)	0.54	0.00 (土 0.00)	0.00	3043.89 (土 15.87)	6.48
5.80	4.94	34.40 (土 1.62)	0.17	231.60 (土 30.46)	1.14	36.45 (土60.40)	0.18	1142.52 (土83.51)	5.64
5.90	2.91	102.26 (土 3.65)	0.30	858.61 (土225.66)	2.50	11.94 (土49.54)	0.03	4483.33 (土272.99)	13.5
6.00	1.73	91.12 (土 23.50)	0.16	270.28 (土 25.63)	0.47	206.27 (土5.51)	0.36	2170.50 (土168.73)	3.75
6.00	3.59	89.19 (土 1.87)	0.32	1149.55 (土90.07)	4.13	41.24 (土13.39)	0.15	4480.31 (土312.27)	16.08
6.30	3.00	85.69 (土 1.61)	0.26	885.50 (土 34.07)	2.66	91.36 (土 33.28)	0.27	1928.18 (土 187.36)	5.78
6.40	4.04	31.15 (土 1.22)	0.13	308.68 (土31.87)	1.25	7.57 (土 34.02)	0.03	1772.28 (土135.50)	7.16
6.50	5.59	60.56 (土 13.97)	0.34	331.47 (土5.63)	1.85	61.45 (土53.93)	0.34	2040.64 (土250.12)	11.41
6.70	3.86	85.09 (土 0.09)	0.33	835.22 (土39.26)	3.22	24.78 (土 3.28)	0.10	4193.81 (土 84.00)	16.19
7.00	8.81	44.73 (土 6.13)	0.39	1119.41 (土21.88)	9.86	22.03 (土19.29)	0.19	2543.45 (土233.71)	22.23
7.30	3.48	104.39 (土 2.18)	0.36	915.82 (土35.99)	3.19	11.92 (土5.12)	0.04	4635.66 (土813.92)	16.13
7.70	8.72	50.23 (土 2.55)	0.44	395.64 (土 3.12)	3.45	21.60 (土45.87)	0.19	1691.77 (土 185.39)	14.75

Table 2 Toxic trace metals (Hg, Cd, Pb and As) concentration and total amount of particular trace metals present in *S. variolaris* at Beruwala reef (n=13) with \pm SD

,	,								
Test diameter	Total gonad	Hg		g		РЬ		As	
(cm)	weight (g)	Average (Hg) (µg/kg)	Total Hg (µg)	Average (Cd) (µg/kg)	Total Cd (µg)	Average (Pb) (µg/kg)	Total Pb (μg)	Average (As) (µg/kg)	Total As (μg)
4.10	0.75	0.00 (土 0.00)	0.00	83.60 (土1.61)	0.06	0.00 (土 0.00)	0.00	(00.0 (0.00
4.10	0.80	188.12 (土 24.29)	0.15	289.72 (土18.01)	0.23	415.03 (土 30.92)	0.33	1866.37 (土325.00)	1.49
4.80	1.12	164.31 (土 7.12)	0.18	201.39 (土35.36)	0.23	772.27 (土211.65)	0.86	3050.80 (土145.32)	3.42
5.00	1.46	92.92 (土6.75)	0.14	247.80 (土33.81)	0.36	309.80 (土 18.40)	0.45	2785.16 (土168.17)	4.07
5.70	3.31	51.51 (土2.21)	0.17	796.99 (土19.78)	2.64	93.70 (土11.10)	0.31	1574.48 (土97.97)	5.21
6.50	7.71	12.90 (土 1.80)	0.10	997.91 (土647.59)	7.69	2.15 (土9.97)	0.02	869.85 (土44.62)	6.71
6.50	8.73	56.61 (土8.09)	0.49	559.31 (土 19.37)	4.88	42.93 (土 26.77)	0.37	1621.43 (土37.61)	14.16
7.20	11.80	6.27 (土4.70)	0.07	379.09 (土3.39)	4.47	00.00 (土 0.00)	0.00	3037.02 (土 34.40)	35.84
7.30	3.66	115.49 (土1.98)	0.42	4590.83 (土155.96)	16.80	87.68 (土 16.02)	0.32	4219.97 (土104.78)	15.45
7.70	6.61	11.87 (土0.13)	0.08	500.57 (土8.29)	3.31	0.00 (土 0.00)	0.00	1696.57 (土198.02)	11.21
8.30	10.72	65.66 (土 1.38)	0.70	1499.39 (土 25.00)	16.07	118.34 (土 35.89)	1.27	4563.32 (土175.59)	48.92

2
Ξ
Ī
2
ų
00
2
Ě
2
2
Ĥ
t
ų.
Ĵ
2
è
S
v
2.
ŧ
ā
ğ
c
-
÷
2
2
č
÷
ž
Ë
Ş.
1
ŝ
٣
÷
Ì
2
2
-
ž
÷
ξ
a
2
÷
2
ŝ
9
ę
č
Y
7
È
0
à
τ
C
ζ
5
<u> </u>
+
0
2
ž
1
į.
5





(See figure on previous page.)

Fig. 3 Relationships of toxic heavy metal concentration in *S. variolaris* at Beruwala reef. **a** Hg versus test diameter, **b** Hg versus total gonad weight, **c** Cd versus test diameter, **d** Cd versus total gonad weight, **e** Pb versus test diameter, **f** Pb versus total gonad weight, **g** As versus test diameter, **h** As versus total gonad weight

Hg levels in *S. variolaris* individuals from Tangalle reef (n=11) ranged from 0.00 (± 0.00) to 188.12 (± 24.29) µg/kg and total level in gonads varied from 0.00 to 0.70 µg. Cd levels ranged from 83.60 (± 1.61) to 4590.83 (± 155.96) µg/kg and total level in gonads varied from 0.06 to 16.07 µg. Pb ranged from 0.00 (± 0.00) to 772.27 (± 211.65) µg/kg and total levels in gonads varied from 0.00 to 1.27 µg. In addition, As in selected individuals ranged from 0.00 (± 0.00) to 4563.32 (± 175.59) µg/kg and total levels in gonads varied from 0.00 to 1.27 µg. In addition, As in selected individuals ranged from 0.00 (± 0.00) to 4563.32 (± 175.59) µg/kg and total levels in gonads varied from 0.00 to 48.92 µg (Table 3).

Relationships with test diameter and gonad weight for trace metal accumulation

The results revealed that there is a positive correlation with between gonadal Hg levels and the test diameter (r=0.826, df=4, p=0.043) and also a between gonad weight (r=0.856, df=4, p=0.030) (Fig. 2a, b) for *S. variolaris* population in Mount-Lavinia. Gonadal Cd levels also had a significant relationship with test diameter (r=0.928, df=4, p=0.008), but strong positive *r* and have significant relationship with gonad weight (r=0.966, df=4, p=0.002) (Fig. 2c, d). Total Pb levels showed no correlation with the test diameter and gonad weight (Fig. 2e, f). Arsenic levels showed a significant positive relationship with test diameter (r=0.820, df=4, p=0.046) and also with the gonad weight (r=0.880, df=4, p=0.021) (Fig. 2g, h).

Hg levels of the Beruwala population showed a significant positive relationship with test diameter (r=0.839, df=11, p<0.001) and gonad weight (r=0.718, df=11, p=0.006) (Fig. 3a, b). Similarly, Cd levels also showed a significant positive relationship with test diameter (r=0.567, df=11, p=0.043) and gonad weight (r=0.709, df=11, p=0.007) (Fig. 3c, d). Pb has a small positive r value and has no significant relationship with test diameter (r=0.267, df=11, p=0.378) and gonad weight (r=0.291, df=11, p=0.334) (Fig. 3e, f). Levels of the trace metal As also had a significant positive relationship with test the diameter (r=0.686, df=11, p=0.010)

Page 9 of 12

and as well as with the gonad weight (r=0.658, df=11, p=0.015) (Fig. 3g, h).

The analysis of the Tangalle population revealed that there is a significant positive relationship between Hg levels and the test diameter (r=0.542, df=9, p=0.085) and with gonad weight (r=0.414, df=9, p=0.205) (Fig. 4a, b). Similarly, cadmium levels showed a strong significant positive relationship with the test diameter (r=0.764, df=9, p=0.006) as well as with the gonad weight (r=0.509, df=9, p=0.110) (Fig. 4c, d). Levels of As also had a significant positive relationship with test diameter (r=0.779, df=9, p=0.005) and gonad weight (r=0.825, df=9, p=0.002) (Fig. 4e, f). Similarly, Pb levels also did not show a significant relationship with test diameter (r=0.151, df=9, p=0.657) and gonad weight (r=0.048, df=9, p=0.888) (Fig. 4g, h).

Discussion

There were low levels of trace metals in the gonads of S. variolaris in all the studied locations. The significant finding was that there were undetectable levels of Pb in S. variolaris gonads collected from the Mount-Lavinia reef. Further, at this location S variolaris population had accumulated As (0.00 ppm-1.81 mg/kg) at a high level (Table 4) compared to other studied metals (Hg and Cd) though this did not exceeded the European Union (EU) (Table 4) and Sri Lanka permitted level. S. variolaris populations in the Beruwala reef was not highly contaminated by Hg (0.03-0.10 mg/kg) and Pb (0.00-0.21 mg/ kg), but highly contaminated by As (1.14-4.64 mg/kg) and Cd (0.15–1.15 mg/kg), which exceeds the permitted level of the 1.00–2.00 mg/kg (Table 4). According to the EU, there were very low levels of Hg (0.00-0.12 mg/kg) in the S. variolaris population of Tangalle reef. However, there were higher levels of Cd (0.08-4.60 mg/kg) and As (0.00-4.563 mg/kg) in this population though they were below the permitted levels. Accumulated Pb levels in S. variolaris (0.00-0.77 mg/kg) varied between permitted level (0.30-1.50 mg/kg) (Table 4). The low levels of toxic trace metals could be seen in the Mount-Lavinia

(See figure on next page.)

Fig. 4 Relationships of toxic heavy metal concentration in *S. variolaris* at Tangalle reef. **a** Hg versus test diameter, **b** Hg versus total gonad weight, **c** Cd versus test diameter, **d** Cd versus total gonad weight, **e** Pb versus test diameter, **f** Pb versus total gonad weight, **g** As versus test diameter, **h** As versus total gonad weight



Table 4 Maximum level of toxic trace metals contaminants of selected seafoods species according to EU and Sri Lanka regulations [22–24]

Contaminants	Food types	Maximum level of contaminants/(mg/kg wet weight)
Pb	Muscle meat of fish	0.30
	Crustaceans	0.50
	Bivalve, molluscs	1.50
Cd	Muscle meat of fish	0.05
	Crustaceans	0.50
	Bivalve, molluscs	1.00
Hg	Fish products and muscle meat	1.00
As	Fish, crustacea, molluscs	1.00-2.00

population when compared to the other populations. The levels of accumulated trace metals in sea urchin gonads represent the environmental condition and thus can be used as bio-indicator to determine the trace metal pollution of that particular coast. As a consequence, most sea urchin species are used as bio-indicators in the world (*Paracentrotus lividus*) [20, 23]. However, there is currently no literature available for trace metal contamination in sea urchins, specifically for Sri Lanka,.

The average levels of accumulated toxic trace metals in sea urchins were in the order from As > Cd > Pb > Hgin Beruwala population, Cd > As > Pb > Hg in Tanagalle population and As > Cd > Hg > Pb in Mount-lavinia population. According to the above order, both Beruwala and Tangalle reefs had the lowest Hg contamination than the Mount-Lavinia sample.

The accumulated levels of all of the tested toxic trace metals (Hg, Cd and As) significantly correlated with the test diameter of S. variolaris population in the Beruwala reef. This indicates that the accumulation of Hg, Cd and As increases with a size of the test. Nevertheless, accumulated Pb levels did not significantly correlate with average test diameter and total gonad weight. Therefore, it seems that the accumulation of Pb does not depend on the total gonad weight or the average test diameter. In the Tangalle sample, Cd and As metals showed a significant relationship between these toxic trace metal accumulations with test diameter while Hg and Pb metals did not significantly correlate with test diameter. However, total gonad weight of S. variolaris showed that accumulation of As significantly correlated though Hg, Cd and Pb did not. The Mount-Lavinia samples showed that there was a significant relationship between toxic trace metals accumulation with average test diameter with Hg, Cd and As metals and Pb has not significantly correlated with average test diameter. Similarly, the total gonad weight of *S. variolaris* revealed a significant relationship with Hg, Cd and As metals and Pb did not because Pb accumulation was not at detectable level. Therefore, it could not be seen any relationship between average test diameter and total gonad weight of Pb accumulation. Finally, As was the only metal which showed a clear significant relationship between both average test diameter and total gonad weight in all the selected sites.

These relationships of *S. variolaris* revealed that toxic trace metals accumulation increase with their age. This would be due to long time accumulation as a result of exposure to the polluted environment in the lifetime. In addition, accumulation of excess trace metals in gonads of sea urchins showed a relationship with their age and a similar result was also reported by Mostafa and Collins [20], because the maximum value of trace metal concentration resulted in the high average test diameter recorded individuals than the low average test diameter recorded individuals. The study conducted by Mostafa and Collins [20] revealed that accumulation of Cd and Pb showed the negative linear relationship between their wet weight and accumulated concentration in the species *P. lividus*.

The frequency distribution of shell diameter and weight of *S. variolaris* range from a minimum of 3.30 to a maximum of 8.90 cm and weight from a minimum of 30.15 g to a maximum of 346.56 g for all the population selected from Sri Lanka. But Smith and Kroh [7] recorded *S. variolaris* having maximum shell diameter of 11 cm from the Visakhapatnam Coast (India). However, 8.90 cm is the highest value recorded for the Sri Lankan *S. variolaris*.

Conclusion

The findings from this study will be helpful to the get an idea about the trace metal accumulation *S. variolaris.* Most importantly, it also indicates that the accumulated levels of trace metals did not exceed the harmful level for human consumption. Hence, due to safe levels of trace metals in *S. variolaris* gonads, there is possibility to export the gonads and gonad products from Sri Lanka, which will be a highly profitable industry in the future.

Abbreviations

IPHT: Institute of Post-Harvest Technology; NARA: National Aquatic Resources Research and Development Agency; Hg: mercury; Pb: lead; Cd: cadmium; As: arsenic; AAS: atomic absorption spectrophotometer; $Pd(NO_3)_2$: palladium nitrate; ppb: parts per billion; EU: European Union.

Authors' contributions

All authors contributed to the manuscript improvement. All authors read and approved the final manuscript.

Author details

¹ Department of Bioprocess Technology, Faculty of Technology, Rajarata University of Sri Lanka, Mihintale, Sri Lanka. ² Department of Physical Sciences, Faculty of Applied Sciences, Rajarata University of Sri Lanka, Mihintale, Sri Lanka. ³ Institute of Post-Harvest Technology (IPHT), National Aquatic Resources Research and Development Agency (NARA), Colombo, Sri Lanka.

Acknowledgements

The authors are grateful to the National Aquatic Resources Research and Development Agency (NARA). Also, we thank to K. Ukuwela for his invaluable comments on manuscript.

Competing interests

The authors declare that they have no competing interests.

Availability of data and materials

The data sets generated and analysed during the current study are available from the corresponding author on reasonable request.

Consent for publication

Not applicable.

Ethics approval and consent to participate

Not applicable.

Funding

This study was supported in part by a grant from the National Aquatic Resources Research and Development Agency (NARA), Crow Island, Colombo 15, Sri Lanka.

Publisher's Note

Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Received: 20 February 2018 Accepted: 7 August 2018 Published online: 16 August 2018

References

- Jakimska A, Konieczka P, Skóra K, Namieśnik J. Bioaccumulation of metals in tissues of marine animals, part I: the role and impact of heavy metals on organisms. Poli J Environ Stud. 2011;20(5):1117–25.
- Hernández OD, Gutiérrez ÁJ, González-Weller D, Lozano G, Melón EG, Rubio C, Hardisson A. Accumulation of toxic metals (Pb and Cd) in the sea urchin *Diadema aff. antillarum* Philippi, 1845, in an oceanic island (Tenerife, Canary Islands). Environ Toxicol. 2010;25(3):227–33.
- De Zoysa HKS, Jinadasa BKKK, Edirisinghe EMRKB. Black sea urchin (Stomopneustes variolaris): nutritional composition and trace metals accumulation of edible sea urchin of Sri Lanka. In: Baker R, editor. LAP LAMBERT Academic Publishing; 2016.
- Jinadasa BKKK, De Zoysa HKS, Jayasinghe GDTM, Edirisinghe EMRKB. Determination of the biometrical parameters, biochemical composition and essential trace metals of edible sea urchin (*Stomopneustes variolaris*) in Sri Lanka. Cogent Food Agric. 2016;2(1):1–12.
- González SJ, Cáceres WC, Ojeda FP. Feeding and nutritional ecology of the edible sea urchin *Loxechinus albus* in the northern Chilean coast. Rev Chil Hist Nat. 2008;81:575–84.
- Reynolds JA, Wilen JE. The sea urchin fishery: harvesting, processing and the market. Mar Resour Econ. 2000;15:115–26.
- Smith AB, Kroh A. The Echinoid Directory Natural History Museum in United Kingdom: World Wide Web electronic publication 2011 [cited 2014-12-06]. http://www.nhm.ac.uk/research-curation/projects/echin oid-directory.
- Kato S, Schroeter SC. Biology of the red sea urchin, Strongylocentrotus franciscanus, and its fishery in California. Mar Fish Rev. 1985;47(3):1–20.

- Morse BL, Hunt HL. Effect of unidirectional water currents on displacement behaviour of the green sea urchin *Strongylocentrous droebachiensis*. J Mar Biol Assoc UK. 2013;93(7):1923–8.
- 10. Siikavuopio SI. Green sea urchin (*Strongylocentrotus droebachiensis*, Müller) in aquaculture: the effects of environmental factors on gonad growth. University of Tromsø, Faculty of Biosciences, Fisheries and Economics, Department of Arctic and Marine Biology; 2009. https://www.resea rchgate.net/profile/Sten_Siikavuopio2/publication/40272651_Green _sea_urchin_Strongylocentrotus_droebachiensis_Muller_in_aquac ulture_the_effects_of_environmental_factors_on_gonad_growth/links/58e256a3a6fdcc41bf991006/Green-sea-urchin-Strongylocentro tus-droebachiensis-Mueller-in-aquaculture-the-effects-of-environmen tal-factors-on-gonad-growth.pdf.
- Kroh A. Stomopneustes variolaris (Lamarck, 1816). In: Kroh A, Mooi R, editors. Retrieved 12-06-2014, from Accessed through: Kroh, A. & Mooi, R. (2014) World Echinoidea Database at http://www.marinespecies.org/ echinoidea/aphia.php?p=taxdetails&id=212440 (2014).
- Giese AC, Krishnaswamy S, Vasu BS, Lawrence J. Reproductive and biochemical studies on a sea urchin, *Stomopneustes variolaris* from Madras Harbor. Comp Biochem Physiol. 1964;13(4):367–80.
- 13. James DB. Ecology of intertidal echinoderms of the Indian seas. J Mar Biol Assoc India. 1982;24(1& 2):124–9.
- Jayakody S. Provisional checklist of sea urchins (Echinodermata: Echinoidea) of Sri Lanka. In: Weerakoon DK, Wijesundara S, editors. The Nat red list 2012 of Sri Lanka. Conservation Status of the Fauna and Flora. Colombo, Sri Lanka: Ministry of Environment; 2012. p. 370–2.
- 15. James DB. Indian echinoderms their resources biodiversity zoogeography and conservation. Glim Aqu Biol. 2008;7:120–32.
- James DB. Research on Indian echinoderms—a review. J Mar Biol Assoc India. 1983;25(1 & 2):91–108.
- James P, Siikavuopio S. A guide to the sea urchin reproductive cycle and staging sea urchin gonad samples; 2011. p. 1–20. https://www.nofim a.no/filearchive/guide-to-sea-urchins_lowres.pdf.
- Salon NA. Echinoderm fisheries of the world: a review. Rotterdam: A A Balkema; 1985. p. 109–24.
- Scheibling RE, Mladenov PV. The decline of the sea urchin, *Tripneustes ventricosus*, fishery of barbados: a survey of fishermen and consumers. Mar Fish Rev. 1987;49(3):62–9.
- Mostafa HM, Collins KJ. Heavy metal concentrations in sea urchin tissues from Egypt, Ireland and United Kingdom. Chem Ecol. 1995;10(1–2):181–90.
- Dincer T, Cakli S. Chemical composition and biometrical measurements of the Turkish Sea Urchin (*Paracentrotus lividus*, Lamarck, 1816). Crit Rev Food Sci Nutr. 2007;47(1):21–6.
- 22. European Commission. Commission Regulation (EC) No 1881/2006 of 19 December 2006 setting maximum levels for certain contaminants in foodstuffs. Off J Eur Union. 2006;L364:5–24.
- Sumner J. Hazards affecting Australian seafood. In: Part 2 supporting information. Adelaide, Australia: SafeFish and the Australian Seafood Cooperative Research Centre; 2011. p. 46.
- Jinadasa BKKK, Jörundsdóttir HÓ, Gunnlaugsdóttir H. Preparation of a standard operation procedure for validation of laboratory methods for trace metal analysis in seafood for national aquatic resources research and development agency (NARA), Sri Lanka. 2010.

Ready to submit your research? Choose BMC and benefit from:

- fast, convenient online submission
- thorough peer review by experienced researchers in your field
- rapid publication on acceptance
- support for research data, including large and complex data types
- gold Open Access which fosters wider collaboration and increased citations
- maximum visibility for your research: over 100M website views per year

At BMC, research is always in progress.

Learn more biomedcentral.com/submissions

